ly for Na, because of Na loss in the beam and low detection efficiency. Furthermore, additional statistical analytical errors are usually introduced when analyzing small precipitates because the overall count rates are much lower than for coarse areas.

- The vector treatment of amphibole compositions is discussed by J. B. Thompson, Jr., in *Characterization of Metamorphism through Mineral Equilibria*, J. M. Ferry, Ed. (Mineralogical Society of America Washington DC, 1982) pp. 1–32.
- America, Washington, DC, 1982), pp. 1–32.
  J. C. Rabbitt [*Am. Mineral.* 33, 263 (1948)] published two anthophyllite compositions showing 0.50 and 0.49 Ca pfu (analyses #15 and 41); W. A. Deer, R. A. Howie, and J. Zussman [*Rock-Forming Minerals, vol. 2, Chain Silicates* (Longmans, Green, London, 1963), pp. 211–229] reported an anthophyllite with 0.492 Ca pfu (analysis #1); and G. N. Starkov [*Zap. Vses. Mineral. Obhch.* 101, 349 (1972)] published an anthophyllite analysis containing 0.569 Ca pfu.
  P. Robinson, in *Pyroxenes, C. T. Prewitt, Ed.*
- P. Robinson, in *Pyroxenes*, C. T. Prewitt, Ed. (Mineralogical Society of America, Washington, DC, 1980), pp. 419–494.
- 16. W. A. Deer, R. A. Howie, J. Zussman, Rock-

Forming Minerals, vol. 2A, Single-Chain Silicates (Wiley, New York, 1978), pp. 41, Table 4, analysis #5.

- 17. P. Robinson, M. Ross, H. Jaffe, *Am. Mineral.* 56, 1005 (1971).
- J. C. Schumacher, personal communication.
   P. J. Treloar and A. Putnis, *Mineral. Mag.* 45, 55
- (1982).
  20. D. L. Kohlstedt and J. B. Vander Sande, in *Electron Microscopy in Mineralogy*, H.-R. Wenk *et al.*, Eds. (Springer-Verlag, Berlin, 1976), pp. 234–237.
- 21. We thank J. Schumacher for supplying a chip of orthoamphibole from locality 6A9 and for providing unpublished microprobe analyses. We also thank P. J. Heaney and M. W. Nyman for reviewing the manuscript. The electron microscopy was undertaken at the transmission electron microscope facility in the Department of Earth and Planetary Sciences at The Johns Hopkins University, Baltimore, MD, supported by National Science Foundation grant EAR89-03603 and instrumentation grant EAR83-00365.

21 May 1992; accepted 20 July 1992

## Responses to Elevated Carbon Dioxide in Artificial Tropical Ecosystems

### Christian Körner\* and John A. Arnone III

Carbon, nutrient, and water balance as well as key plant and soil processes were simultaneously monitored for humid tropical plant communities treated with  $CO_2$ -enriched atmospheres. Despite vigorous growth, no significant differences in stand biomass (of both the understory and overstory), leaf area index, nitrogen or water consumption, or leaf stomatal behavior were detected between ambient and elevated  $CO_2$  treatments. Major responses under elevated  $CO_2$  included massive starch accumulation in the tops of canopies, increased fine-root production, and a doubling of  $CO_2$  evolution from the soil. Stimulated rhizosphere activity was accompanied by increased loss of soil carbon and increased mineral nutrient leaching. This study points at the inadequacy of scaling-up from physiological baselines to ecosystems without accounting for interactions among components, and it emphasizes the urgent need for whole-system experimental approaches in global-change research.

**G**reater carbon sequestering by terrestrial ecosystems, increasing amounts of leaf area per unit of land area, reduced water consumption, and greater efficiency of mineral nutrient capture are some of the common predictions for plant and ecosystem behavior in a future world with high levels of  $CO_2$ (1–5). We evaluated these assumptions in an experiment with closed ecosystems of tropical rain forest plants of various life forms arranged in complex communities.

The only available data about in situ  $CO_2$  responses of a natural terrestrial ecosystem, Alaskan tundra (6), revealed no such growth stimulation, whereas semiaquatic salt-marsh vegetation, whose nutritional situation is comparable to agricultural conditions, not surprisingly showed positive biomass responses similar to those found in crops (7, 8). Forests, which make

up about 90% of the global biomass, have not yet been studied under artificially enhanced CO<sub>2</sub> levels, and current predictions of their behavior in a CO2-rich world are based on data largely derived from experiments with isolated seedlings (4, 9). Before success is realized from long-term and largescale CO<sub>2</sub> enrichment experiments on natural forests (10), investigations with experimental ecosystems provide an excellent means for the assessment of CO<sub>2</sub> effects in highly structured ecosystems. We selected a humid tropical ecosystem because this biome represents about 40% of the global biomass and because plant responses to  $CO_2$ are predicted to be more pronounced under high temperatures (11).

We constructed four such model ecosystems by enclosing identically structured populations of 15 tropical plant species (12) in 17-m<sup>3</sup> polyethylene-covered houses. Plants in each house shared a common soil volume of 1.3 m<sup>3</sup> and ground area of 6.7 m<sup>2</sup>. The two to six individuals of each species in each

SCIENCE • VOL. 257 • 18 SEPTEMBER 1992

ecosystem were located in exactly the same position in the four houses. Soils consisted of a mixture of silicate sand and vermiculite, overlain by a thin layer of leaf and bark compost, which was supplemented with 20 g per square meter of timed-release fertilizer pellets. Houses were situated in a climateconditioned greenhouse at the University of Basel, Switzerland, and received natural daylight. Closed air-circulation systems changed the air in each house 11 times per hour and included activated charcoal filters and dew point and temperature controllers with dehumidifying water traps. The experimental ecosystems were not designed to mimic a specific natural situation but to represent model stands with a characteristic vertical stratification.

All ecosystems were allowed to stabilize under low CO<sub>2</sub> levels (340  $\mu$ l of CO<sub>2</sub> per liter of air) for 30 days before the experiment. No differences in stand structure {for example, leaf area index [LAI; area of leaves (m<sup>2</sup>) per square meter of ground] and mean height of individual plants} were apparent among houses after this stabilization period. Daytime CO<sub>2</sub> was maintained at 340  $\mu$ l of CO<sub>2</sub> per liter in two of the houses, while the other two received 610  $\mu$ l of CO<sub>2</sub> per liter during the experiment. The experiment was terminated after 3 months, 1 month after the LAI stabilized at a value typical for humid tropical forests (13).

Total ecosystem biomass at both CO<sub>2</sub> levels more than doubled within 3 months (from about 700 to 1500 g of dry mass per square meter); thus, the experimental conditions supported vigorous growth. Despite this increase, neither the biomass of the entire ecosystem nor that of the individual species (14) responded significantly to elevated  $CO_2$  (Table 1), although a slight positive effect was detected. This result is in contrast to that of Ziska et al. (15), who reported dramatic stimulation of growth in tropical plants isolated in pots that contained rich soil. However, in the long run, marginal gains, such as those found here under elevated CO<sub>2</sub>, could still result in more rapid maturation of stands.

One of the key determinants of carbon fixation by ecosystems is the amount of photosynthetic machinery per unit area of land, commonly expressed as the LAI. Elevated  $CO_2$  has been shown to enable plants to achieve greater relative carbon gains under low irradiances (4, 16, 17), and therefore greater leaf retention on plants in the shade would be expected, leading to greater LAIs (18). The LAI in our study, measured with an electronic canopy analyzer (19), increased linearly in all ecosystems during the first 60 days from 3.4 to approximately 7, after which it leveled off (Fig. 1). Steady-state LAI was accompanied by the onset of leaf litter production, which was

Department of Botany, University of Basel, Schönbeinstrasse 6, CH-4056 Basel, Switzerland.

<sup>\*</sup>To whom correspondence should be addressed.

**Table 1.** Effect of elevated CO<sub>2</sub> levels on final biomass and litter production (measured in grams per square meter). Mean  $\pm$  SE of two houses for each CO<sub>2</sub> level (ambient and elevated CO<sub>2</sub>) are shown. With the use of the two-tailed *t* test, *P* gives the probability that the results from the two CO<sub>2</sub> treatments are not the same. Biomass partitioning is represented as a percent of the total system biomass which is calculated as the mean of two houses. The biomass of roots  $\leq 1$  mm in diameter was measured for the entire soil depth with the use of mesh in-growth cylinders. Air circulation systems were regularly switched between ambient and elevated CO<sub>2</sub> houses to minimize the potential effects of systematic errors.

Component	Treatment					Mean		
	Ambient CO <sub>2</sub>			Elevated CO <sub>2</sub>			differ- ence (%	P
		ount of mass	% of total biomass		ount of mass	% of total biomass	change in actual biomass)	Ρ
Biomass (total)	1408	± 5	100.0	1558	± 52	100.0	+11	0.1022
Aboveground	1165	± 7	82.7	1241	± 51	79.6	+ 6	0.2741
Leaf	391	± 7	27.7	413	± 6	26.5	+ 5	0.1371
Stem	774	± 14	55.0	828	± 45	53.1	+ 7	0.3658
Roots	243	± 2	17.3	318	± 2	20.4	+30	0.0011
Diameter ≤1 mm	32	± 5	2.3	52	± 3	3.3	+63	0.0658
Diameter 1 to 3 mm	103	± 3	7.3	141	± 2	9.1	+37	0.0068
Diameter >3 mm	108	± 1	7.7	125	± 6	8.0	+16	0.1113
Litter production	8.4	4 ± 0.0		11.6	6 ± 0.0		+38	0.0002

greater under elevated  $CO_2$  (Table 1). Final LAI and light extinction (1 to 2% reaching the ground) (Fig. 2) were similar to those reported for closed forest canopies in Puerto Rico and Malaysia (13).

Leaves in ecosystems exposed to elevated  $CO_2$  levels contained twice as much total nonstructural carbohydrates (TNC; 18% versus 8% of dry mass) as those grown under ambient  $CO_2$  levels. Thus, TNC-free leaf biomass under elevated  $CO_2$  levels was actually 6% less (not significant) than that of leaves under ambient  $CO_2$  levels. TNC accumulation increased dramatically with height in the canopy, and up to one-third of total leaf dry matter was composed of TNC in plants exposed to high levels of  $CO_2$  (Fig. 2). Monocots tended to accumulate less than dicots even when leaves from the same canopy layer were compared. No evidence of carbohydrate dissipation from



**Fig. 1.** Leaf canopy development under ambient and elevated  $CO_2$  levels. The LAI is measured in area of leaves (m<sup>2</sup>) per square meter of ground.

sun-exposed to shaded leaves of the same plant was observed. Accumulation of TNC under elevated  $CO_2$  levels has been also observed in a number of crop plants (20, 21), but this vertical stratification has not been described. In the houses with elevated  $CO_2$ , top-canopy leaves that contained the largest amounts of starch were also more yellow in *Ficus benjamina* and *Tetrastigma voinerianum* than in counterparts in the houses with ambient  $CO_2$  levels. This observation suggests that increased  $CO_2$  levels under full illumination can have an adverse effect, as has been shown to occur in some potted crop plants under higher  $CO_2$  concentrations [1000 µl of  $CO_2$  per liter (22, 23)] than used here.

There are several reasons that carbohydrates might overflow in CO2-fertilized leaves (24): (i) reduced activity of carbon sinks, (ii) restriction of assimilate transport, or (iii) limitation of phosphorylation in the chloroplasts. Factors that have been shown to result in some of these conditions, which lead to increased TNC accumulation in leaves, are nutrient shortage (21) and the restriction of root growth by pot volume (25) and by soil penetration resistance (26). Because plants in our experiment grew in a large, common volume of soil unlikely to restrict root growth, the next logical step was to evaluate the extent to which plants were limited in the amount of available nutrients (nitrogen in particular).

Total nutrient pool size remained constant because nutrients lost in drainage water were recycled (nitrogen losses from denitrification were assumed to be negligible compared to the size of the nitrogen pool). Nitrogen uptake by plants was the same at both  $CO_2$  levels, although more nitrogen was allocated to roots and less to leaves under elevated  $CO_2$  levels (Table 2). Similar observations have been reported for temperate, forest-tree species (27).

Although total biomass more than doubled during the experiment for both control



Fig. 2. Increases in leaf nonstructural carbohydrate (TNC) with increasing height in the canopy (A) corresponding to irradiance at the end of the experiment (B). Side light was shaded with green plastic slats around each house.

SCIENCE • VOL. 257 • 18 SEPTEMBER 1992

and elevated  $CO_2$  treatments, one could argue that soil nutrient limitations may have precluded a significant biomass response to the elevated  $CO_2$ . However, this does not seem to have been the case because ions of all major nutrients were abundant in percolation water after all watering in which the soil was saturated (Table 3). Amounts of nitrate, potassium, and phosphate in percolation water of ecosystems with high  $CO_2$  levels were twice as high as those in ecosystems with low  $CO_2$  levels. Hence, although mineral nutrients were not supplied to permit luxury consumption, they were available at sufficient levels to have allowed for a  $CO_2$  stimulation of growth.

Another possible explanation for the unexpectedly small stimulation of biomass production under elevated CO<sub>2</sub> levels could

**Table 2.** Effect of elevated CO<sub>2</sub> levels on final system nitrogen distribution [grams of N per square meter of ground (g N m<sup>-2</sup>); mean  $\pm$  SE of two houses for each CO<sub>2</sub> level (ambient and elevated) are shown; *P* is as in Table 1. The CHN analyzer was used for plant tissue, and Kjeldahl procedures were used for total soil N]. Starting soil nitrogen for all houses was 39.9  $\pm$  0.2 g N m<sup>-2</sup>, which included 2.9 g N m<sup>-2</sup> in 3-month time-release Osmocote pellets (Sierra, Heerlen, the Netherlands).

0	Treat	Mean			
Component	Ambient CO <sub>2</sub> Elevated CO <sub>2</sub>		difference (%)	Р	
N in biomass (total)	10.47 ± 0.02	10.46 ± 0.09	0	0.8914	
Aboveground	8.76 ± 0.11	8.18 ± 0.09	- 7	0.0551	
Leaf	$5.54 \pm 0.03$	$5.05 \pm 0.02$	- 9	0.0054	
Stem	$3.22 \pm 0.14$	$3.13 \pm 0.07$	- 3	0.6234	
Roots	$1.71 \pm 0.09$	2.28 ± 0.01	+33	0.0245	
Diameter ≤1 mm	$0.45 \pm 0.05$	$0.68 \pm 0.05$	+51	0.0021	
Diameter 1 to 3 mm	$0.73 \pm 0.02$	$1.00 \pm 0.01$	+37	0.0034	
Diameter >3 mm	0.53 ± 0.01	$0.60 \pm 0.05$	+13	0.2939	
Litter	$0.09 \pm 0.00$	$0.12 \pm 0.00$	+33	0.0002	
Soil (total)	$34.68 \pm 0.24$	$34.43 \pm 0.01$	- 1	0.4125	

**Table 3.** Effect of elevated CO<sub>2</sub> levels on nutrients (measured in milligrams per square meter) in ecosystem percolation water over the course of the experiment sampled after soil-saturating "rains"; N, nitrogen; P, phosphorus. Mean  $\pm$  SE of two houses for each CO<sub>2</sub> level (ambient and elevated) are shown; *P* is as in Table 1. Analysis methods: ammonium (NH<sub>4</sub><sup>+</sup>), Indophenol method (Berthelot's reaction); nitrate (NO<sub>3</sub><sup>-</sup>), ion chromatography; phosphate (PO<sub>4</sub><sup>-3</sup>), molybdenum blue; K<sup>+</sup>, flame photometry; Mg<sup>2+</sup>, atomic absorption.

Nutrients	Trea	tment	Mean difference	Ρ
	Ambient CO <sub>2</sub>	Elevated ÇO <sub>2</sub>	(%)	
NH4 <sup>+</sup> -N	0.78 ± 0.59	0.37 ± 0.12	- 53	0.5090
NO <sub>3</sub> <sup>4</sup> N PO <sub>4</sub> <sup>3</sup> -P K <sup>+</sup>	$1.58 \pm 0.40$	$3.78 \pm 0.93$	+139	0.0551
PO <sub>4</sub> <sup>−3</sup> -P	$16.59 \pm 4.60$	$37.00 \pm 8.64$	+123	0.0636
K+ <sup>‡</sup>	$7.62 \pm 2.03$	20.39 ± 6.31	+168	0.0831
Mg <sup>2+</sup>	$1.34 \pm 0.61$	$1.34 \pm 0.64$	0	0.9993

**Table 4.** Effect of elevated  $CO_2$  levels on below-ground carbon at the end of the experiment. Mean  $\pm$  SE of two houses for each  $CO_2$  level (ambient and elevated) are shown; *P* is as in Table 1. Methods: fine-root length, line intersection method (samples taken from in-growth cylinders); soil  $CO_2$  efflux, mobile cuvettes and infrared gas analysis; soil carbon of pooled samples from each house, dichromate oxidation.

Company	Trea	Mean	P	
Component	Ambient CO <sub>2</sub>	Elevated CO <sub>2</sub>	difference (%)	Ρ
Fine-root length				
(meters per square meter of ground)	0150 . 000	5004 . 450		0.0000
Diameter ≤ 1 mm	3152 ± 629	$5304 \pm 158$	+ 68	0.0800
Diameter >1 mm	380 ± 37	461 ± 68	+ 21	0.4058
Soil CO <sub>2</sub> efflux				
$(\mu mol of CO_2 m^{-2} \text{ ground s}^{-1})$	$1.9 \pm 0.2$	$3.6 \pm 0.2$	+ 89	0.0176
Soil carbon content (g C $m^{-2}$ )				
Start				
Absolute loss in 94 days	$75 \pm 10$	$300 \pm 35$	+300	0.0261
% loss of starting C in 94 days	$5.9 \pm 0.6$	$23.6 \pm 3.3$	+300	0.0346

SCIENCE • VOL. 257 • 18 SEPTEMBER 1992

be a downward adjustment of photosynthesis (20, 25, 28). Actual canopy  $CO_2$  uptake measured at the end of the experiment was almost twice as high in houses with elevated  $CO_2$  levels as in those with ambient  $CO_2$  levels. However, uptake rates of canopies in both  $CO_2$  treatments, after an accounting for  $CO_2$  evolution from the soil, were identical (7.5  $\mu$ mol of  $CO_2$  per square meter of ground per second) when  $CO_2$ -rich ecosystems were exposed to ambient  $CO_2$  for short periods. This result indicated that no downward adjustment had occurred. Also, leaf diffusive conductances and stomatal densities did not differ between treatments (14).

A large portion of the assimilates produced in excess of what occurred as starch in canopies in the houses with high  $CO_2$ levels appeared to have been allocated to sinks below the ground level. Fine-root biomass (Table 1) was significantly greater at elevated CO<sub>2</sub> levels, a frequently reported observation (29, 30), and  $CO_2$  efflux from soils almost doubled (Table 4). Efflux rates measured in houses with ambient  $CO_2$ levels (1.9  $\mu$ mol of CO<sub>2</sub> m<sup>-2</sup> ground s<sup>-1</sup>) were similar to the mean (2.0  $\mu$ mol of CO<sub>2</sub> m<sup>-2</sup> ground s<sup>-1</sup>) calculated for annual rates for 13 different tropical forests reported by Raich and Nadelhoffer (31). Greater fineroot mass would account for some of the increase in efflux found in our ecosystems at high  $CO_2$  levels, but much of the  $CO_2$  must have arisen from stimulated microbial activity that was a result of increased fine-root turnover or root exudation (32), as indicated by one-third greater soil xylanase and protease activity in houses with high  $CO_2$ levels (33). Increased turnover of fresh root tissue (for example, release of xylose) and exudation of soluble substrate (carbohydrates and amino acids) possibly stimulated the breakdown of soil organic matter and would explain why cumulative soil carbon losses were three times greater at the end of the experiment under elevated CO<sub>2</sub> levels (Table 4). Elliott et al. (34) also reported CO<sub>2</sub>-stimulated rates of soil CO<sub>2</sub> efflux in shortgrass prairie microcosms. The average daily consumption of water by the experimental ecosystems did not differ between CO<sub>2</sub> treatments. This substantiates the lack of differences observed in LAI and leaf diffusive conductance between the different CO2 treatments and agrees with data of Reekie and Bazzaz (35).

We find it alarming that  $CO_2$  fertilization can promote losses of soil carbon and the release of mineral nutrients similar to the effects achieved when sugar is added to soils (36). These effects may lead to longlasting depletion of mineral nutrients if unbound ions are leached from the rhizosphere. There is a danger that such processes may be masked in the short term, because elevated  $CO_2$  has been shown to stimulate growth, at least temporarily, even under serious nutrient shortage (29, 37).

Another concern raised by our results is that the deleterious levels of starch produced in leaves in the top of canopies under elevated CO<sub>2</sub> levels may cause alterations in dominance relations in plant communities as CO2 rises. Accumulation of TNC appears to be a result of a specific local dissipation problem at the cellular or chloroplast level because we see no obvious reason for reduced sink activity or longdistance transport problems in the ecosystems with high  $CO_2$  levels. This was indicated by significant increases in the allocation of carbon to the rhizosphere (Tables 1 and 4) and a lack of TNC concentration gradients in stem tissue in plants growing under elevated  $CO_2$  (14).

Finally, our results challenge thinking about the biological effects of a doubling of atmospheric CO<sub>2</sub> levels. Observed over long periods, natural ecosystems are believed to have a nearly balanced carbon budget, with photosynthetic uptake equal to respiratory release of  $CO_2$  (38). As a result, the ecosystem CO<sub>2</sub> compensation point should be close to ambient  $CO_2$ levels, as was found in the arctic tundra (39). It is uncertain how rapidly ecosystem CO<sub>2</sub> compensation points will adjust to rising CO<sub>2</sub> levels. If ecosystem compensation points have not tracked the rapid increases in atmospheric CO2 levels over the past 100 years, present ambient levels may already represent a massive  $CO_2$  overload to some ecosystems. Our data indicate that elevated CO2 may not necessarily lead to greater carbon sequestering by terrestrial ecosystems, but that it will likely result in greater carbon turnover.

#### **REFERENCES AND NOTES**

- 1. F. I. Woodward, G. B. Thompson, I. F. McKee, Ann. Bot. 67, 23 (1991).
- 2. B. R. Strain and J. D. Cure, Eds., Direct Effects of Increasing Carbon Dioxide on Vegetation (Publ. ER-0238, U.S. Department of Energy, Washington, DC, 1985)
- 3. F. A. Bazzaz and E. D. Fajer, Sci. Am. 266, 68 (1992). P. G. Jarvis, Philos. Trans. R. Soc. London Ser. B 4.
- 324, 369 (1989).
- 5. H. A. Mooney, B. G. Drake, R. J. Luxmoore, W. C. Oechel, L. F. Pitelka, Bioscience 41, 96 (1991). 6. N. E. Grulke et al., Oecologia 83, 485 (1990).
- J. D. Cure, in (2), pp. 99-116.
- P. S. Curtis, B. G. Drake, P. W. Leadley, W. J. Arp, D. F. Whigham, *Oecologia* 78, 20 (1989).
- 9. D. Eamus and P. G. Jarvis, Adv. Ecol. Res. 19, 1 (1989).
- 10. C. Körner, in Vegetation Dynamics and Global Change, H. Shugart and A. Solomon, Eds. (Chapman & Hall, New York, in press).
- 11. S. P. Long, Plant Cell Environ. 14, 729 (1991).
- 12. The 43 plants per house making up the experimental ecosystems were the tree and shrub species Cecropia peltata (Cecropiaceae), Ficus benjamina (Moraceae), Theobroma cacao (Sterculiaceae), and Piper tiliaefolium (Piperaceae); the climbing vines Monstera dubia (Araceae), Piper nigrum (Piperaceae), and Tetrastigma voineri-

anum (Vitaceae); the ground creepers Ficus pumila (Moraceae) and Philodendron scandens (Araceae); and the herbaceous monocots Hedv chium coccineum, Elettaria cardamomum (Zingiberaceae), Ctenanthe lubbersiana (Marantaceae), Heliconia humilis (Musaceae), Areca lutescens (Arecaceae), and Aechmea mexicana (Bromeliaceae), Except for T. cacao, which was raised from seed, all plants of the same species were propagules of the same clone.

- T. Kira and K. Yoda, in Tropical Rain Forest 13 Ecosystems, H. Lieth and M. J. A. Werger, Eds. (Elsevier, Amsterdam, 1989), pp. 55–71. C. Körner and J. A. Arnone III, data not shown.
- 15. L. H. Ziska et al., Oecologia 86, 383 (1991).
- 16. R. M. Gifford, Aust. J. Plant Physiol. 4, 99 (1977). F. A. Bazzaz, J. S. Coleman, S. R. Morse, Can. J. 17. For. Res. 20, 1479 (1989). 18. I. Nijs et al., Planta 177, 312 (1989).
- The LAI was measured with a Licor LI-2000 Can-19. opy Analyzer (electronic fisheye sensor) (Lincoln, Nebraska) calibrated at the harvest against actual leaf areas measured by a photoplanimeter. LI-2000 readings were corrected for background values, and the data presented here are corrected and calibrated values.
- 20 T. W. Sasek, E. H. DeLucia, B. R. Strain, Plant Physiol. Suppl. 75 (abstr. 26) (1984).
- S.-C. Wong, Photosynth. Res. 23, 171 (1990) 21 22. R. D. Wulff and B. R. Strain, Can. J. Bot. 60, 1084 (1982)
- 23 T. W. Sasek, E. H. DeLucia, B. R. Strain, Plant Physiol. 78, 619 (1985).
- 24. M. Stitt, Plant Cell Environ. 14, 741 (1991). R. B. Thomas and B. R. Strain, Plant Physiol. 96, 25 627 (1991)
- J. Masl, G. D. Farquhar, R. M. Gifford, Aust. J. 26 Plant Physiol. 17, 465 (1990).
- 27. R. J. Norby, E. G. O'Neill, R. J. Luxmoore, Plant

Physiol. 82, 83 (1986)

- W. J. Arp, Plant Cell Environ. 14, 869 (1991) 28 R. J. Norby and E. G. O'Neill, New Phytol. 117, 29.
- 515 (1991). 30. H. Poorter, S. Pot, H. Lambers, Physiol. Plant. 73,
- 553 (1988). 31. J. W. Raich and K. J. Nadelhoffer, Ecology 70,
- 1346 (1989) 32. R. J. Norby, R. J. Luxmoore, E. G. O'Neill, D. G.
- Weller, Environmental Sciences Division Publication 2318 (Oak Ridge National Laboratory, Oak Ridge, TN, 1984).
- 33. Soil xylanase and protease activities were determined by F. Schinner (personal communication).
- 34. E. T. Elliott, H. W. Hunt, J. K. Detling, J. C. Moore, D. E. Reuss, Bull. Ecol. Soc. Am. 71, 147 (1990)
- 35. E. G. Reekie and F. A. Bazzaz, Oecologia 79, 212 (1989).
- J. P. E. Anderson and K. H. Domsch, Soil Biol. 36. Biochem. 10, 215 (1978).
- 37. R. J. Luxmoore, R. J. Norby, E. G. O'Neill, in Forest Plants and Forest Protection, 18th International Union of Forestry Research Organizations (IUFRO), World Congress, Div. 2, 1987) (IUFRO Secretariate, Vienna, 1987), vol. 1, pp. 178–183.
  P. G. Jarvis, *Tree Physiol.* 2, 347 (1986).
  D. T. Tissue and W. C. Oechel, *Ecology* 68, 401
- (1987).
- We thank our colleagues in the Department of 40. Botany for their assistance during this study. especially W. Hilti for quantifying canopy devel opment, M. Würth for the determination of TNC, W. Flückiger and S. Braun for chemical analysis of soils and water, and E. Schreier for artwork. H. Lambers, R. Norby, B. Schmid, and M. Diemer provided helpful comments on an earlier version of this manuscript

27 April 1992; accepted 9 July 1992

# The Present Is Not the Key to the Past: A Polar Forest from the Permian of Antarctica

### Edith L. Taylor,\* Thomas N. Taylor, N. Rubén Cúneo

An in situ Upper Permian fossil forest in the central Transantarctic Mountains near the Beardmore Glacier includes 15 permineralized trunks in growth position; the paleolatitude of the site was approximately 80° to 85° south. Numerous leaves of the seed fern Glossopteris are present in the shale in which the trunks are rooted. The trunks are permineralized and tree rings reveal that the forest was a rapidly growing and young forest, persisting in an equable, strongly seasonal climate—a scenario that does not fit with some climate reconstructions for this time period.

**F**ossil forests represent an important data source that can be utilized to reconstruct past climates. High-latitude fossil forests are rare, with instances known from Cretaceous deposits on Alexander Island (Antarctic peninsula) (1) and Paleocene-Eocene sediments on Axel Heiberg Island (Canadian Arctic) (2). These sites provide an opportunity to examine forest density and productivity at high-latitude sites in the past. The preservation of the wood also allows for an analysis of tree ring data.

E. L. Taylor and T. N. Taylor, Byrd Polar Research Center and Department of Plant Biology, Ohio State University, Columbus, OH 43210.

\*To whom correspondence should be addressed.

When combined with data from compression-impression floras, vegetation reconstruction is greatly enhanced. This biological input represents an important parameter that can be used to calibrate climate models based primarily on physical data.

During our 1991 to 1992 field season, we examined an in situ fossil forest on a flat bench at the top of the northward extending ridge of Mount Achernar (Fig. 1). The forest is exposed over an area of about 20 by 12 m (84°22'23"S, 164°37'56"E; Buckley Island Quadrangle) (3). Fifteen stumps are preserved in growth position and occur within shaley floodplain deposits (4) in the upper part of the Buckley Formation. Based on palynomorphs recovered at nearby sites (5), the age of this deposit is considered to be Late Permian.

N. R. Cúneo, Museo Paleontológico Egidio Feruglio, 9 de Julio 655 (9100) Trelew, Chubut, Argentina.